Fine particulate air pollution and hospital emergency room visits for respiratory disease in urban areas in Beijing, China, in 2013

Qin Xu
Xia Li
Shuo Wang
Chao Wang
Fangfang Huang

See next page for additional authors

Follow this and additional works at: https://ro.ecu.edu.au/ecuworkspost2013

Part of the Atmospheric Sciences Commons, Environmental Health Commons, and the Other Medical Sciences Commons


This Journal Article is posted at Research Online.
https://ro.ecu.edu.au/ecuworkspost2013/1709
RESEARCH ARTICLE

Fine Particulate Air Pollution and Hospital Emergency Room Visits for Respiratory Disease in Urban Areas in Beijing, China, in 2013

Qin Xu1,2*, Xia Li1,2,3*, Shuo Wang4, Chao Wang1,2, Fangfang Huang1,2, Qi Gao1,2, Lijuan Wu1,2, Lixin Tao1,2, Jin Guo1,2, Wei Wang1,2,5*, Xiuhua Guo1,2*

1 Department of Epidemiology and Health Statistics, School of Public Health, Capital Medical University, Beijing, China, 2 Beijing Municipal Key Laboratory of Clinical Epidemiology, Beijing, China, 3 Graduate Entry Medical School, University of Limerick, Limerick, Ireland, 4 Beijing Chaoyang Hospital, Capital Medical University, Beijing, China, 5 School of Medical Sciences, Edith Cowan University, Perth, WA, Australia

☯ These authors contributed equally to this work.
* statguo@ccmu.edu.cn (XG); wei.wang@ecu.edu.au (WW)

Abstract

Background

Heavy fine particulate matter (PM_{2.5}) air pollution occurs frequently in China. However, epidemiological research on the association between short-term exposure to PM_{2.5} pollution and respiratory disease morbidity is still limited. This study aimed to explore the association between PM_{2.5} pollution and hospital emergency room visits (ERV) for total and cause-specific respiratory diseases in urban areas in Beijing.

Methods

Daily counts of respiratory ERV from Jan 1 to Dec 31, 2013, were obtained from ten general hospitals located in urban areas in Beijing. Concurrently, data on PM_{2.5} were collected from the Beijing Environmental Protection Bureau, including 17 ambient air quality monitoring stations. A generalized-additive model was used to explore the respiratory effects of PM_{2.5}, after controlling for confounding variables. Subgroup analyses were also conducted by age and gender.

Results

A total of 92,464 respiratory emergency visits were recorded during the study period. The mean daily PM_{2.5} concentration was 102.1±73.6 μg/m^3. Every 10 μg/m^3 increase in PM_{2.5} concentration at lag0 was associated with an increase in ERV, as follows: 0.23% for total respiratory disease (95% CI: 0.11%-0.34%), 0.19% for upper respiratory tract infection (URTI) (95% CI: 0.04%-0.35%), 0.34% for lower respiratory tract infection (LRTI) (95% CI: 0.14%-0.53%) and 1.46% for acute exacerbation of chronic obstructive pulmonary disease (AECOPD) (95% CI: 0.13%-2.79%). The strongest association was...
identified between AECOPD and PM$_{2.5}$ concentration at lag0-3 (3.15%, 95%CI: 1.39%-4.91%). The estimated effects were robust after adjusting for SO$_2$, O$_3$, CO and NO$_2$. Females and people 60 years of age and older demonstrated a higher risk of respiratory disease after PM$_{2.5}$ exposure.

**Conclusion**

PM$_{2.5}$ was significantly associated with respiratory ERV, particularly for URTI, LRTI and AECOPD in Beijing. The susceptibility to PM$_{2.5}$ pollution varied by gender and age.

**Introduction**

Because of the rapid industrialization and economic growth experienced over the past ten years, air pollution has become a serious issue in China. Heavy haze-fog episodes have frequently occurred in recent years [1, 2]. From 2013 to 2014, the annual mean fine particulate matter (PM$_{2.5}$, particles with an aerodynamic diameter $\leq$ 2.5 $\mu$m) concentration for most of the 31 provincial capital cities in China exceeded the Chinese Ambient Air Quality Standards Grade II standard (35 $\mu$g/m$^3$) [3]. In particular, as the capital of China, Beijing has experienced a more serious challenge in terms of deteriorating air quality because of the coal burning and wind-blown dust from the surrounding industrial cities as well as the vehicle exhaust from the rapidly growing number of automobiles [4]. For example, starting in early January 2013, Beijing experienced multiple prolonged periods of severe smog; the peak hourly concentration of ambient PM$_{2.5}$ soared to 800 $\mu$g/m$^3$, and the annual ambient PM$_{2.5}$ concentration in 2013 reached 89.5 $\mu$g/m$^3$ [5]. The extremely high ambient PM$_{2.5}$ concentration has attracted extensive public attention because of the adverse health effects.

Toxicological evidence suggests that the exposure of PM$_{2.5}$ can cause lung inflammation and affect pulmonary immune function [6]. During the past two decades, several studies have demonstrated the effects of PM$_{2.5}$ on the mortality for total respiratory disease [5, 7] or on the number of hospital admissions due to cause-specific respiratory diseases, such as pneumonia, chronic obstructive pulmonary disease and asthma [8–11]. However, few reports have directly evaluated the effects of PM$_{2.5}$ on total and cause-specific respiratory diseases using hospital emergency room visits (ERV) as a morbidity indicator. In addition, most of these studies were conducted in Western developed countries; however, because of the different level of effects, such as ambient PM$_{2.5}$ levels, characteristics of ambient PM$_{2.5}$, population susceptibility, and weather patterns, there is still a need to assess the health effects of PM$_{2.5}$ exposure in developing countries such as China.

It was not until March 1, 2012, that the PM$_{2.5}$ level was required to be included in the air quality standards in China [1]. Previous studies evaluating the short-term adverse effects of ambient particles pollution in China have primarily focused on the effects of PM$_{10}$ (particles with an aerodynamic diameter $\leq$ 10 $\mu$m) [12, 13]. Researchers have suggested that PM$_{2.5}$ is more harmful to health than PM$_{10}$ is, because smaller particles are more readily able to deposit deeper into the lungs and can carry larger concentrations of adsorbed or condensed toxic air pollutants per unit mass. Although numerous studies have reported the effects of PM$_{2.5}$ on mortality [14, 15], a recent meta-analysis has noted that the observed association between ambient PM$_{2.5}$ and morbidity in China has not been validated [16]. Although one study reported the association between PM$_{2.5}$ and hospital admissions for respiratory diseases in Beijing, the correlation was explored from the perspective of geography and spatial analysis using...
grey correlation analysis, and it considered particles measuring less than 1.0 μm [17]. Another study employed hospital admissions data from one hospital in the Haidian district in Beijing, but it did not detect the effects of PM$_{2.5}$ on cause-specific respiratory diseases [18].

The availability of the thorough monitoring network for PM$_{2.5}$ from Beijing Environmental Protection Bureau since October 2012 provided us with an opportunity to assess the adverse effects of PM$_{2.5}$ pollution. This study aimed to explore association between PM$_{2.5}$ concentration and ERV for total and cause-specific respiratory diseases using a time-series design by controlling for long-term trends and seasonality in daily ERV, day of the week, and weather variables in Beijing. Subgroup analyses were also performed using different age groups and gender to identify the most susceptible subpopulations.

**Materials and Methods**

**Ethics statement**

The study protocol was approved by the Institutional Review Board at the School of Public Health, Capital Medical University. Written informed consent was not required because we used aggregated data and no individualized data. The patient information was anonymized and de-identified prior to analysis.

**Study site**

Beijing is located in the north of China (latitude 39°54’N and longitude 116°25’E). As the capital, Beijing is the political, economical and cultural center of China, with six urban districts (Dongcheng, Xicheng, Chaoyang, Haidian, Fengtai, Shijingshan) and ten suburban districts (Changping, Mentougou, Daxing, Fangshan, Tongzhou, Shunyi, Huairou, Pinggu, Minyun and Yanqing). According to the Chinese Health Statistical Yearbook, in 2012, Beijing had a population of 20,690,000, with 10,685,164 males (51.64%) and 10,004,836 females (48.36%). Approximately 75% of Beijing’s total population live in six urban districts, which cover all high population density areas in Beijing (> 5,000 people/km$^2$) [19].

**Data collection**

Daily data on emergency room visits from Jan 1 to Dec 31, 2013, were collected from ten general hospitals located in urban areas in Beijing. To obtain a representative sample of the urban population, one to three publicly funded hospitals in each urban district were selected (Fig 1). Patient data were derived from the computerized medical record system including age, gender, visit date and diagnosis at discharge. ERV for respiratory diseases were extracted according to the principal diagnosis, using the International Classification of Diseases-(ICD) 10th Revision Code of J00-J99. Specifically, we focused on total respiratory disease (ICD-10: J00-J99), upper respiratory tract infection (URTI) (ICD-10: J00, J02-J06), lower respiratory tract infection (LRTI) (ICD-10: J12-J22, J44 [except J44.103 and J44.901]), acute exacerbation of chronic obstructive pulmonary disease (AECOPD) (ICD-10: J44.103, J44.901) and asthma (ICD-10: J45-46).

Since October 2012, daily average and hourly real-time air pollution data were released to the public through a web platform (http://zx.bjmemc.com.cn/) by Beijing Environmental Protection Bureau, which established 35 ambient air quality monitoring stations in the 16 districts of Beijing city. Daily 24-hour average concentration data for PM$_{2.5}$ (unit, μg/m$^3$), sulfur dioxide (SO$_2$) (unit, μg/m$^3$), ozone (O$_3$) (unit, μg/m$^3$), nitrogen dioxide (NO$_2$) (unit, μg/m$^3$) and carbon monoxide (CO) (unit, mg/m$^3$) levels during the study period were collected, including data from 17 ambient air quality monitoring stations located in urban areas in Beijing city.
The rates of the missing values from the 17 monitoring stations ranged from 9.0% to 15.6%. Some daily data were missing for all districts, primarily due to technical problems (i.e., website maintenance) during the study period. Imputation will produce errors, therefore, we did not fill in the missing values. We used the mean PM2.5 concentrations from the 17 monitoring stations to represent the daily exposure of the urban population. In addition, to allow for the effects of weather conditions, during the same period, we obtained daily meteorological data from the Chinese Meteorological Bureau, including the mean temperature and relative humidity.

Data analysis

A generalized-additive model (GAM) with the Poisson link function was used to explore the association between the daily PM2.5 concentration and outcomes, after adjusting for temperature, relative humidity, day of the week (DOW), public holidays and influenza outbreaks. First, the basic model was set up, excluding air pollutant variables. The natural cubic smoothing spline (ns) functions of calendar time, daily mean temperature and relative humidity were incorporated in the model, allowing for nonlinear relationships with the outcomes. The partial

(Fig 1). The rates of the missing values from the 17 monitoring stations ranged from 9.0% to 15.6%. Some daily data were missing for all districts, primarily due to technical problems (i.e., website maintenance) during the study period. Imputation will produce errors, therefore, we did not fill in the missing values. We used the mean PM2.5 concentrations from the 17 monitoring stations to represent the daily exposure of the urban population. In addition, to allow for the effects of weather conditions, during the same period, we obtained daily meteorological data from the Chinese Meteorological Bureau, including the mean temperature and relative humidity.

Data analysis

A generalized-additive model (GAM) with the Poisson link function was used to explore the association between the daily PM2.5 concentration and outcomes, after adjusting for temperature, relative humidity, day of the week (DOW), public holidays and influenza outbreaks. First, the basic model was set up, excluding air pollutant variables. The natural cubic smoothing spline (ns) functions of calendar time, daily mean temperature and relative humidity were incorporated in the model, allowing for nonlinear relationships with the outcomes. The partial
autocorrelation function was used to guide the selection of the best degrees of freedom (df) for the spline function of calendar time. Note that the absolute values of the partial autocorrelation function for the model residuals had to be < 0.1 for the first 2 lag days [20], and then the Akaike Information Criterion (AIC) was used to determine the best df for calendar time. Thus, a natural cubic smoothing spline for calendar time (df = 11) was used to control for the seasonal and long-term trends. Natural cubic smoothing spline for temperature and relative humidity (df = 3) on current day of ERV occurred was incorporated into the model, based on previous studies [21, 22]. The day of the week (DOW), public holidays and influenza outbreaks were all categorical variables. After the basic model was established, the air pollutant variables were then added. The final model is described below:

$$\text{Log}[E(Y_t)] = \alpha + \beta_1 \text{PM2.5}_{t-i} + ns(\text{Temp}_0, \text{df} = 3) + ns(\text{RH}_0, \text{df} = 3) + ns(\text{Time}, \text{df} = 11) + \beta_2 \text{DOW} + \beta_3 \text{Holiday} + \beta_4 \text{Influenza}$$

$E(Y_t)$ is the expected count for respiratory ERV on day $t$; $PM2.5_{t-i}$ is the mean PM$_{2.5}$ concentration from 17 monitoring stations on day $t$, and $i$ is the day lag; $ns$ is the natural cubic splines; $\text{Temp}_0$ and $\text{RH}_0$ are the daily mean temperature (°C) and relative humidity (%) of the current day, respectively; $\text{Time}$ denotes long-term trends and seasonality using the calendar time days; $\text{DOW}$ is the day of the week; $\text{holiday}$ indicates a public holiday on day $t$ (0 indicates no holiday, and 1 indicates a holiday); $\text{influenza}$ is a dummy variable for the weeks, with a number of influenza ERV exceeding the 75 percentile in a year [23].

Smoothing function in GAM was used to graphically analyze the exposure-response relation-ship to verify the assumption of linearity between the predicted log-relative risk of respiratory ERV and PM$_{2.5}$ concentration. The linear effects of PM$_{2.5}$ were then estimated for the current day and up to 5 day before the outcome (lag0 to lag5). Considering that a single-day lag model might underestimate the association [19], the overall cumulative effects were estimated using 2-day, 4-day and 6-day moving averages of PM$_{2.5}$ concentrations (lag0-1, lag0-3 and lag0-5). We also investigated whether the associations were still sensitive after adjusting for the other gaseous pollutants (SO$_2$, O$_3$, CO or NO$_2$) in two-pollutant models [24]. In the single-pollutant models, PM$_{2.5}$ was placed in the model alone; in the two-pollutant models, SO$_2$, O$_3$, CO or NO$_2$ was jointly included with PM$_{2.5}$. Effects across age groups (0–14 years, 15–34 years, 35–59 years, and ≥ 60 years) and genders were examined using the respiratory ERV subgroups for the health outcomes to identify the most susceptible subpopulation [25]. A $Z$-test was then used to test the statistical significance of differences by gender or age by calculating 

\[
(\beta_1 - \beta_2) / \sqrt{SE_1^2 + SE_2^2},
\]

where $\beta_1$ and $\beta_2$ are the estimates coefficient for the two categories (i.e., male and female patients), and $SE_1$ and $SE_2$ are the respective standard errors [26].

Sensitivity analyses were conducted to examine the impact of PM$_{2.5}$ on total respiratory ERV using different df: time trend (6–12), temperature (2–6) and relative humidity (2–6). Studies have shown that temperature effects on health would lag for more than 10 days [27]; therefore, the effects of PM$_{2.5}$ were also checked in our model using the 14-day moving average temperature [28]. All sensitivity analyses were performed only for total respiratory ERV in the whole population.

All analyses were performed using the “mgcv” and “nlme” packages in the statistical envi-ronment R 3.2.2 (Development Core Team, 2015). The results are presented as the percentage change (PC) and 95% confidence interval (CI) in the daily respiratory ERV per 10 μg/m$^3$ increase in PM$_{2.5}$ concentration.
Results

Descriptive statistics

A total of 92,464 emergency visits for total respiratory diseases were recorded from the ten hospitals in 2013, with URTI accounting for the largest proportion (55.3%). The mean daily count for total respiratory disease was 253 (range, 129 to 556), and 45.5% were male patients. The mean patient age was 44.1 years for male patients and 43.0 years for female patients. There were 19 patients (0.021%) with missing gender information and 6 patients (0.006%) with missing age information, which had little effect on the subgroup analyses. The mean daily visits for cause-specific respiratory diseases were URTI (113 visits), LRTI (122 visits), AECOPD (2 visits) and asthma (6 visits) (Table 1).

During the study period, the overall mean daily PM$_{2.5}$ concentration was 102.1 $\mu$g/m$^3$ (range, 6.7 $\mu$g/m$^3$ to 508.5 $\mu$g/m$^3$). In terms of the Chinese Ambient Air Quality Standards Grade II standards for 24-hour average PM$_{2.5}$ concentration (≤ 75 $\mu$g/m$^3$), 45.7% (155 days) of the daily PM$_{2.5}$ concentrations were below this standard during the study period. However, in terms of the WHO Air Quality Standards (24-hour average PM$_{2.5}$ concentration ≤ 25 $\mu$g/m$^3$), only 8.8% (30 days) met the standard. The mean daily concentrations of SO$_2$, O$_3$, CO and NO$_2$ were 31.2 $\mu$g/m$^3$, 104.3 $\mu$g/m$^3$, 2.2 mg/m$^3$ and 66.7 $\mu$g/m$^3$, respectively (Table 2). All Spearman correlations among the pollutants and weather variables were statistically significant at 0.05 level (2-tailed). Generally, PM$_{2.5}$ was strongly correlated with SO$_2$, CO and NO$_2$ ($r_s$ = 0.56, 0.80 and 0.72, respectively) and negatively correlated with O$_3$ ($r_s$ = −0.15). The time series graph showed the daily variations of ERV for respiratory diseases and PM$_{2.5}$ concentrations during study period (Fig 2).

Associations between PM$_{2.5}$ and ERV for respiratory disease

There were clear exposure-response relationships between PM$_{2.5}$ concentration and total respiratory ERV (Fig 3). The exposure-response relationships were approximately linear, with a tiny fluctuation when the PM$_{2.5}$ concentrations were below 200 $\mu$g/m$^3$ and a sharper response at higher PM$_{2.5}$ concentrations.

Table 1. Characteristics and distribution of the main health outcomes and sub-categories of the different gender and age groups.

<table>
<thead>
<tr>
<th>Age group</th>
<th>Total respiratory disease</th>
<th>URTI</th>
<th>LRTI</th>
<th>AECOPD</th>
<th>Asthma</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
</tr>
<tr>
<td>Total</td>
<td>92464</td>
<td>241 (207–275)</td>
<td>51134</td>
<td>126 (108–155)</td>
<td>34904</td>
</tr>
<tr>
<td>Gender</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Male</td>
<td>42095</td>
<td>110 (96–127)</td>
<td>22119</td>
<td>55 (47–68)</td>
<td>16877</td>
</tr>
<tr>
<td>Female</td>
<td>50350</td>
<td>130 (109–151)</td>
<td>29003</td>
<td>71 (60–89)</td>
<td>18021</td>
</tr>
<tr>
<td>Age group</td>
<td>Total respiratory disease</td>
<td>URTI</td>
<td>LRTI</td>
<td>AECOPD</td>
<td>Asthma</td>
</tr>
<tr>
<td></td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
<td>N (25-75)</td>
</tr>
<tr>
<td>0–14</td>
<td>5159</td>
<td>13 (9–18)</td>
<td>3002</td>
<td>8 (5–11)</td>
<td>2030</td>
</tr>
<tr>
<td>15–34</td>
<td>38228</td>
<td>99 (83–117)</td>
<td>28304</td>
<td>72 (59–87)</td>
<td>7973</td>
</tr>
<tr>
<td>35–59</td>
<td>24792</td>
<td>63 (54–76)</td>
<td>14111</td>
<td>35 (28–44)</td>
<td>8471</td>
</tr>
<tr>
<td>60–</td>
<td>24279</td>
<td>65 (54–74)</td>
<td>5713</td>
<td>14 (10–19)</td>
<td>16428</td>
</tr>
</tbody>
</table>

Note: URTI-upper respiratory tract infection, LRTI-lower respiratory tract infection, AECOPD-acute exacerbation of chronic obstructive pulmonary disease, N-counts of event, daily median (25th percentile and 75th percentile) (25th percentile and 75th percentile) of daily event counts.

doi:10.1371/journal.pone.0153099.t001
Fig 4 shows the associations between the PM$_{2.5}$ concentration and total respiratory ERV. We observed statistically significant associations between the total respiratory ERV and PM$_{2.5}$ concentration on the current day (lag0), the previous day (lag1), the previous three day (lag3), the 2-day moving average (lag0-1) and the 4-day moving average (lag0-3). We estimated increases of 0.23% (95% CI: 0.11%-0.34%), 0.12% (95% CI: 0.01%-0.22%), 0.17% (95% CI: 0.07%-0.27%), 0.22% (95% CI: 0.09%-0.35%) and 0.21% (95% CI: 0.05%-0.37%) in total respiratory ERV associated with every 10 $\mu$g/m$^3$ increase in PM$_{2.5}$ at lag0, lag1, lag3, lag0-1 and lag0-3, respectively. The largest estimated effect was observed on the current day (lag0). The estimated effects of PM$_{2.5}$ on total respiratory ERV changed little after adjusting for O$_3$, but fluctuated with attenuating tendencies after adjusting for SO$_2$ or CO and increasing tendencies after adjusting for NO$_2$. Significant results were observed at different lag days in the two-pollutant models (Fig 4).

Table 3 lists the effects of PM$_{2.5}$ on cause-specific respiratory ERV. Every 10 $\mu$g/m$^3$ increase of PM$_{2.5}$ on the current day (lag0) was associated with an increase in ERV: 0.19% for URTI (95% CI: 0.04%-0.35%), 0.34% for LRTI (95% CI: 0.34%-0.53%), and 1.46% for AECOPD (95% CI: 0.13%-2.79%), respectively. Single-day delayed effects of PM$_{2.5}$ were significant with URTI on lag1 and lag3, with LRTI on lag2 and lag3, and with AECOPD on lag1 and lag2. For the cumulative effects, the 2-day moving average PM$_{2.5}$ concentration (lag0-1) was significantly associated with ERV for URTI, LRTI and AECOPD. The estimated effect of PM$_{2.5}$ on AECOPD at lag0-3 (3.15%, 95% CI: 1.39%-4.91%) was the largest. No significant association was found between PM$_{2.5}$ and asthma ERV at any lag patterns (P>0.05). The estimated effects of PM$_{2.5}$ on cause-specific respiratory ERV in two-pollutant models are shown in the supporting information, S1 Table.

The association between PM$_{2.5}$ and total respiratory ERV differed by gender and age group (Fig 5). PM$_{2.5}$ exposure resulted in slightly larger effects in females than in males at all lag patterns. Meanwhile, the gender differences were statistically significant at lag2, lag0-3 and lag0-5 (P<0.05). We observed that the estimated effects of PM$_{2.5}$ on total respiratory ERV among people 60 years of age and older were significant at all lag patterns and higher when compared with other age groups. In terms of the estimated effects, the differences between age groups were statistically significant in the following categories: between people 60 years of age and older and the age group younger than 15 years at lag0-1, lag0-3 and lag0-5; between people 60 years of age and older and the age group 15 years of age and older and younger than 35 years at

Table 2. Summary of the environmental and meteorological variables in urban areas in Beijing during the study period.

<table>
<thead>
<tr>
<th>Variables</th>
<th>Mean</th>
<th>SD</th>
<th>Min</th>
<th>25th</th>
<th>Median</th>
<th>75th</th>
<th>Max</th>
</tr>
</thead>
<tbody>
<tr>
<td>PM$_{2.5}$ ($\mu$g/m$^3$)</td>
<td>102.1</td>
<td>73.6</td>
<td>6.7</td>
<td>53.1</td>
<td>82.4</td>
<td>129.5</td>
<td>508.5</td>
</tr>
<tr>
<td>SO$_2$ ($\mu$g/m$^3$)</td>
<td>31.2</td>
<td>25.7</td>
<td>3.6</td>
<td>11.0</td>
<td>23.3</td>
<td>43.7</td>
<td>133.8</td>
</tr>
<tr>
<td>O$_3$ ($\mu$g/m$^3$)</td>
<td>104.3</td>
<td>68.3</td>
<td>4.4</td>
<td>53.0</td>
<td>89.4</td>
<td>146.5</td>
<td>305.4</td>
</tr>
<tr>
<td>CO (mg/m$^3$)</td>
<td>2.2</td>
<td>1.5</td>
<td>0.3</td>
<td>1.2</td>
<td>1.9</td>
<td>2.6</td>
<td>17.7</td>
</tr>
<tr>
<td>NO$_2$ ($\mu$g/m$^3$)</td>
<td>66.7</td>
<td>25.0</td>
<td>12.4</td>
<td>50.1</td>
<td>62.3</td>
<td>78.3</td>
<td>171.5</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>11.3</td>
<td>11.6</td>
<td>-12.6</td>
<td>1.2</td>
<td>11.0</td>
<td>22.2</td>
<td>29.0</td>
</tr>
<tr>
<td>Relative humidity (%)</td>
<td>58.7</td>
<td>17.3</td>
<td>18.9</td>
<td>46.7</td>
<td>59.0</td>
<td>73.3</td>
<td>93.3</td>
</tr>
</tbody>
</table>

Abbreviations: PM$_{2.5}$-particles with an aerodynamic diameter ≤ 2.5 μm, SO$_2$-sulfur dioxide, O$_3$-ozone, CO-carbon monoxide, NO$_2$-nitrogen dioxide, SD-standard deviation, Max-maximum, and Min-minimum.

doi:10.1371/journal.pone.0153099.t002
Fig 2. Time series plot of emergency room visits for respiratory diseases (number of daily cases) and PM$_{2.5}$ concentrations in Beijing, China during study period. Abbreviations: URTI-upper respiratory tract infection, LRTI-lower respiratory tract infection, AECOPD-acute exacerbation of chronic obstructive pulmonary disease, PM$_{2.5}$-particles with an aerodynamic diameter $\leq$ 2.5 μm.

doi:10.1371/journal.pone.0153099.g002
Sensitivity analysis results

S2 Table shows the percentage change in total respiratory ERV associated with a 10 μg/m³ increase in PM$_{2.5}$ on the current day (lag$_0$), under different df for calendar time, temperature and relative humidity. The AIC of the basic model was lowest, and the estimated effect became stable when df of spline for calendar time was 11; different df of spline for temperature and relative humidity had little effect on the estimated effect, suggesting that the findings on the associations of PM$_{2.5}$ concentration with respiratory ERV were robust in our study. The estimated effects changed little when the 14-day moving average temperature was controlled for in our model. As shown in S3 Table, similar results have been found.

Discussion

This study provided strong evidence of the associations between PM$_{2.5}$ and respiratory disease morbidity in Beijing. PM$_{2.5}$ was significantly associated with ERV for total respiratory disease and cause-specific respiratory diseases, including URTI, LRTI, and AECOPD. Females and people 60 years of age and older suffered a higher risk from PM$_{2.5}$ exposure. To our knowledge, this study is one of the few that have thoroughly investigated the short-term effects of PM$_{2.5}$ air pollution on total and cause-specific respiratory diseases.
Previous studies have identified significant positive associations between PM$_{2.5}$ and respiratory morbidity in Beijing, but they were restricted to one or two districts [17, 18]. To obtain more representative results for the population in the entire city, we correlated the daily ERV for respiratory diseases from ten hospitals located in the six urban districts in Beijing with the citywide daily mean PM$_{2.5}$ concentration from 17 monitoring stations. We were not able to

![Figure 4. Percentage increase of total respiratory ERV associated with a 10 μg/m$^3$ increase in PM$_{2.5}$ concentrations using the single- and two-pollutant models.](doi:10.1371/journal.pone.0153099.g004)

Table 3. Percentage changes with 95% CI in cause-specific respiratory ERV associated with a 10 μg/m$^3$ increase in PM$_{2.5}$ concentrations for different lag structures.*

<table>
<thead>
<tr>
<th>Lag Days</th>
<th>URTI</th>
<th>LRTI</th>
<th>AECOPD</th>
<th>Asthma</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>PC</td>
<td>95% CI</td>
<td>PC</td>
<td>95% CI</td>
</tr>
<tr>
<td>lag0</td>
<td>0.19</td>
<td>(0.04, 0.35)</td>
<td>0.34</td>
<td>(0.14, 0.53)</td>
</tr>
<tr>
<td>lag1</td>
<td>0.14</td>
<td>(0.00, 0.27)</td>
<td>0.08</td>
<td>(-0.09, 0.25)</td>
</tr>
<tr>
<td>lag2</td>
<td>-0.09</td>
<td>(-0.22, 0.05)</td>
<td>0.19</td>
<td>(0.03, 0.36)</td>
</tr>
<tr>
<td>lag3</td>
<td>0.14</td>
<td>(0.00, 0.27)</td>
<td>0.19</td>
<td>(0.02, 0.35)</td>
</tr>
<tr>
<td>lag4</td>
<td>0.02</td>
<td>(-0.12, 0.16)</td>
<td>0.09</td>
<td>(-0.08, 0.26)</td>
</tr>
<tr>
<td>lag5</td>
<td>0.02</td>
<td>(-0.12, 0.16)</td>
<td>-0.10</td>
<td>(-0.27, 0.08)</td>
</tr>
<tr>
<td>lag0-1</td>
<td>0.20</td>
<td>(0.03, 0.37)</td>
<td>0.27</td>
<td>(0.06, 0.48)</td>
</tr>
<tr>
<td>lag0-3</td>
<td>0.18</td>
<td>(-0.03, 0.39)</td>
<td>0.24</td>
<td>(-0.02, 0.50)</td>
</tr>
<tr>
<td>lag0-5</td>
<td>0.10</td>
<td>(-0.17, 0.37)</td>
<td>0.21</td>
<td>(-0.11, 0.53)</td>
</tr>
</tbody>
</table>

*statistically significant results at the 5% level are indicated in bold.

Abbreviations: URTI-upper respiratory tract infection, LRTI-lower respiratory tract infection, AECOPD-acute exacerbation of chronic obstructive pulmonary disease, PC-percentage change.

doi:10.1371/journal.pone.0153099.t003
Fig 5. Percentage change with 95%CI in total respiratory ERV associated with a 10 μg/m³ increase in PM$_{2.5}$ concentrations by gender and age. Note: The $P$ value obtained from the Z-test for the difference between the two relative risks derived from the subgroup analysis. $P_1$: 60 years of age and older vs the age group younger than 15 years; $P_2$: 60 years of age and older vs the age group 15 years of age and older and younger than 35 years; $P_3$: age 60 years of age and older vs the age group 35 years of age and older and younger than 60 years.

doi:10.1371/journal.pone.0153099.g005
allot each patient a unique PM$_{2.5}$ concentration because we did not have individual information on his or her residential address. We assumed that residents would be more likely to go to the nearest hospital in Beijing, therefore, we assigned the monitored concentrations to patients based on their proximity to the hospitals, and then we used generalized additive mixed model (GAMM) to examine the impacts of the spatially assigned PM$_{2.5}$ concentration on health outcomes. However, the estimated effects using spatially assigned PM$_{2.5}$ concentrations in the GAMM were slightly smaller, and the confidence intervals were larger than those using mean PM$_{2.5}$ in the GAM (supporting information, S1 Fig and S4 Table), which was not consistent with the results reported by Xu M, et al [21], likely due to the larger standard error generated by more missing values of air pollutants in the GAMM in our study or because of the smaller spatial variation of PM$_{2.5}$ concentrations in urban areas compared to that in the entire city.

Toxicological studies have proposed that acute impairment of the lung cellular defense and immune system might be causative mechanisms for the short-term effects of PM$_{2.5}$ on respiratory disease [29]. Our current findings on the short-term associations between PM$_{2.5}$ and respiratory ERV are consistent with other toxicological and epidemiological findings. To date, many epidemiological studies have examined and reported the adverse health effects of PM$_{2.5}$ on hospital admissions for LRTI, COPD and asthma [8–11]; however, few reports have directly addressed total respiratory disease using ERV as a morbidity indicator. Our study found statistically significant associations between PM$_{2.5}$ and ERV for total respiratory disease and cause-specific diseases, including URTI, LRTI and AECOPD. Likely due to the higher PM$_{2.5}$ concentration in Beijing in 2013, a longer lag association was observed in our study compared to previous reports of a 0–2 day lag [22, 30]. We found that the effects of PM$_{2.5}$ on LRTI were higher overall compared to the effects on URTI in our study. LRTI, including pneumonia and acute bronchitis, is more severe and can cause a higher disease burden during an individual’s life span than URTI; therefore, most previous studies likely focused on the effect of PM$_{2.5}$ on LRTI rather than URTI [31, 32]. A study conducted in Rome showed a significant association between PM$_{2.5}$ and LRTI at lag$_2$, with a higher effect (2.82%, 95%CI: 0.52%-5.19%) than the estimated effect in our study (0.19%, 95%CI: 0.09%-0.36%) [30]. Studies of the effects of PM$_{2.5}$ on COPD admissions are rare and inconsistent [9, 31, 32]. We found that the effects of PM$_{2.5}$ on AECOPD are the largest compared with other respiratory diseases groups. Due to the slow progression and chronic nature of the disease, COPD can also result in a massive and growing disease burden that affects a patient’s health quality. Thus, COPD patients should stay at home as long as possible during a PM$_{2.5}$ pollution outbreak to avoid possible acute exacerbation. We did not find any statistically significant associations between PM$_{2.5}$ and asthma ERV, likely due to the small daily number of asthma patients. A simulation study suggested that the power of a time-series study of the acute health effects of air pollution can be increased by increasing either the mean daily count of the outcome or the time-series length [33]. In our study, the total cases of AECOPD and asthma were small, which limited our calculation power. In addition, air pollution health effects over a short time interval (e.g., 1-year time series study) are often small; and lower estimated effects also decrease power [33]. The lack of power may inhibit the identification of an association with specific causes of respiratory diseases that have low counts, as is the case with asthma, especially when the estimated effect is low [34].

When examining the associations between PM$_{2.5}$ and total respiratory ERV, we considered the confounding factors of the other gaseous pollutants. The estimated effects of PM$_{2.5}$ on total respiratory disease were affected by the inclusion of SO$_2$, CO or NO$_2$, but not O$_3$ in the two-pollutant models, possibly caused by the strong correlations between PM$_{2.5}$ and SO$_2$ ($r_s = 0.56$), PM$_{2.5}$ and CO ($r_s = 0.80$) or PM$_{2.5}$ and NO$_2$ ($r_s = 0.72$). In studies that have estimated the adverse effects of PM$_{2.5}$ in multi-pollutant models, positive associations between PM$_{2.5}$ and
respiratory causes persisted after adjusting for other gaseous pollutants [8, 24, 35]. In Taiwan, the PM$_{2.5}$ concentration on the current day (lag0) increased hospital admissions for COPD on cool days (<25°C), and the risk estimates reduced after adjustment for O$_3$, CO or NO$_2$, and increased after adjustment for SO$_2$ in the model [8]. In a time-series study conducted in New York City, PM$_{2.5}$ was shown to increase emergency department visits for asthma during warm season in single pollutant models, but the risk estimate was attenuated once SO$_2$, O$_3$, CO or NO$_2$ was included in the model [35]. In another time-series study conducted in Hong Kong, the risk estimate of PM$_{2.5}$ on COPD emergency hospital admissions also became unstable after adjusting for CO and NO$_2$ [24]. These results may reflect the actual difference in the toxicity of the corresponding air pollutants themselves, but it is unlikely to differentiate the independent effects of each air pollutant in multi-pollutant models [21]. In the present study, PM$_{2.5}$ was positively associated with respiratory causes ERV in both single-pollutant and multi-pollutant models at different lags, reflecting the actual adverse effects of PM$_{2.5}$ on human respiratory system.

The shape of the exposure—response relationship between PM$_{2.5}$ concentration and respiratory disease risk is crucial to determine the pattern of the adverse response. Most studies have assumed linear associations between particulate matter air pollution and health outcomes in epidemiological studies [7, 12, 18, 19, 24]. Schwartz et al. explored the exposure—response association between PM$_{2.5}$ and daily deaths by using a hierarchical model in the United States, and found a linear relationship without threshold [36]. This result is support by a cohort study conducted in Europe, which indicated PM$_{2.5}$ was associated with natural-cause mortality, even within concentration ranges well below the present European annual mean limit value (<25 μg/m$^3$) [37]. Tian L et al found that in Hong Kong, the exposure-response curve between PM$_{2.5}$ and COPD was almost identical to that noted by Schwartz et al [24]. However, using regression splines to model the exposure-response relationship between PM$_{2.5}$ and ischemic heart disease morbidity in Beijing, Xie W et al showed that the curve was non-linear, with predicted effect values less than zero at lower concentrations (0–90 μg/m$^3$), a stable response at intermediate concentrations (90–180 μg/m$^3$) and then a relatively shallower response at higher PM$_{2.5}$ concentrations, which is partly consistent with our results [19]. Samoli et al provided evidence for deviation from linearity at lower level of particulate distribution in a multicity study, which estimated the exposure—response relationships between PM$_{10}$ and mortality [38]. They also proposed that the different city-specific relations can be explained partly by factors characterizing the air pollution mix and climate [38]. For example, Wong CM et al found no clear exposure—response relationships in warm seasons between respiratory mortality and PM$_{10}$, but observed a positive exposure—response relationship when PM$_{10}$ concentrations up to 80 μg/m$^3$ in the cool seasons [39].

We found that the effects of PM$_{2.5}$ pollution on total respiratory ERV were significantly greater in females than in males. In previous studies, the gender-specific short-term effects of air pollution were inconsistent [25, 40, 41]. Kan H, et al suggested that nonsmokers may be more sensitive to PM$_{2.5}$ exposure [25], whereas the prevalence of smoking in females is lower than that in males in the Chinese population [42]. Moreover, Sunyer, et al suggested that different particles deposition patterns between males and females may partly explain these differences [43]. The age subgroup analysis results suggested that elderly people might be more susceptible to PM$_{2.5}$ exposure. Elderly people may have a weaker immune system and a higher prevalence of chronic respiratory diseases, thus, they are more vulnerable to PM$_{2.5}$ pollution. Generally, in accordance with previous studies [17, 25, 44], our study suggested that special attention should be paid to elderly people, in terms of their PM$_{2.5}$ exposure.

Compared with previous studies focusing on the associations between PM$_{2.5}$ pollution and respiratory diseases in relatively lower polluted areas in developed countries, our study
investigated the association between PM$_{2.5}$ pollution and respiratory ERV in a heavy polluted city (i.e., Beijing, China). Given the high level of PM$_{2.5}$ pollution in Beijing, we were able to explore the exposure-response relationship in a wide range of PM$_{2.5}$ concentrations. We used respiratory ERV as the health outcomes, and these unscheduled respiratory diseases therapy were more likely to be community-acquired and might reflect the acute effects of PM$_{2.5}$ pollution. Compared to the other study in Beijing [18], our air pollution data were entirely derived from the 17 monitoring stations in urban areas in Beijing, which were more representative of the general population exposure than data from a single monitoring station.

Several limitations of our study should be mentioned. First, exposure measurement error was inevitable in this time-series study, which has been shown to bias estimates downward [45]. Second, we could not identify the re-admissions for patients with the same respiratory disease from the available data. However, it was possible that some patients, particularly children and elderly people, were admitted to hospitals multiple times in a short time period. Such repeated admissions could lead to temporal dependence among the outcomes, which may lead to an underestimation of the variance of PM$_{2.5}$ pollution risk estimates [46]. Third, we only collected 1 year of emergency data from ten general hospitals in the current study due to the constraints of emergency data availability in multiple hospitals, which might have led to bias in the results and also limit our research power. Therefore, further studies are warranted.

Conclusions

In conclusion, our study provides evidence that PM$_{2.5}$ pollution can increase the risk of respiratory ERV, particularly for URTI, LRTI and AECOPD in urban areas in Beijing. The strongest association was between AECOPD and the 4-day moving average PM$_{2.5}$ concentration. Different ages and genders varied in their susceptibility to PM$_{2.5}$ pollution. The government should pay more attention to the health effects of PM$_{2.5}$ pollution and provide appropriate solutions to control PM$_{2.5}$ pollution to improve air quality.

Supporting Information

S1 Database. Air pollution data, meteorological data and emergency data. (XLS)

S1 Fig. Percentage increase of total respiratory ERV associated with a 10 $\mu$g/m$^3$ increase in PM$_{2.5}$ concentrations in GAM and GAMM. (TIF)

S1 Table. Percentage changes with 95% CI in cause-specific respiratory ERV associated with a 10 $\mu$g/m$^3$ increase in PM$_{2.5}$ concentrations for the different lag structures in two-pollutant models. * (DOC)

S2 Table. Percentage changes with 95% CI in total respiratory ERV associated with a 10 $\mu$g/m$^3$ increase in the PM$_{2.5}$ concentrations at lag0 by different degrees of freedom (df) for time, temperature, and relative humidity. (DOC)

S3 Table. Percentage changes with 95% CI in total respiratory ERV associated with a 10 $\mu$g/m$^3$ increase in PM$_{2.5}$ concentrations when the current day’s temperature and 14-day moving average temperature were controlled. (DOC)
S4 Table. Percentage changes with 95% CI in cause-specific respiratory ERV associated with a 10 $\mu$g/m$^3$ increase in PM$_{2.5}$ concentrations for different lag structures in single-pollutant GAMM.

Acknowledgments
The authors thank the doctors from the emergency departments in each hospital for their assistance with emergency data collection.

Author Contributions
Conceived and designed the experiments: XG WW. Performed the experiments: QX XL WW XG. Analyzed the data: QX XL. Contributed reagents/materials/analysis tools: QX XL SW CW FH QG LW LT JG. Wrote the paper: QX XL SW CW FH QG LW LT JG WW XG.

References


